



Perspective

Perspective: Why might removing carnivores maintain or increase risks for domestic animals?

L. Mark Elbroch^{a,*,1}, Adrian Treves^{b,1}^a Panthera, 8 West 40th Street, 18th Floor, NY 10018, USA^b University of Wisconsin-Madison, 550 North Park Street, Madison, WI 53706, USA

ARTICLE INFO

Keywords:

Carnivore
Lethal removal
Livestock conflict
Wildlife management

ABSTRACT

Human-carnivore conflict is still characterized by lethal control, even while some evidence suggests that carnivore removal may not affect the likelihood of future livestock predation, or that it may even exacerbate the problem. Here we propose five non-exclusive, and likely additive, hypotheses for why lethal removals could fail to mitigate livestock-carnivore conflict. We also propose a methodological change in the scale of conflict analyses from populations to smaller social networks, and encourage public education that includes discussions about the potential consequences for communities with livestock following the killing of carnivores, in addition to broader outreach about both the costs and benefits of living with carnivores.

1. Introduction

The management of human-carnivore conflict is still characterized by the lethal removal of carnivores (Treves and Karanth, 2003; Lorand et al., 2022), in part because such actions are seen as addressing rather than ignoring the problem, regardless of whether they reduce further conflict (Naughton-Treves et al., 2003; Linnell, 2011; Dickman et al., 2013). State wildlife agencies in the USA, for example, increase regional hunting quotas in response to livestock losses: “Several factors can trigger a target area [increased carnivore hunting], including the number of cougars being killed for livestock damage or public safety concerns in the area (Oregon Department of Fish and Wildlife, 2019).” A growing body of evidence, however, suggests that removing carnivores—whether via general culling or targeted removal—may not address the real causes of carnivore-livestock conflict (e.g., cheetah, *Acinonyx jubatus*, social behaviors; Melzheimer et al., 2020), or that removals may even exacerbate the problem by leaving the culprits in place or disrupting carnivore social networks (Haber, 1996; Woodroffe et al., 2006; Santiago-Avila et al., 2018; Natrass et al., 2020). Several recent studies of pumas (*Puma concolor*), gray wolves (*Canis lupus*), black-backed jackals (*Canis mesomelas*) and caracals (*Caracal caracal*), among other carnivores, suggest that killing one or more carnivores can increase risks for remaining livestock (Peebles et al., 2013; Minnie et al., 2016; Teichman et al., 2016; Santiago-Avila et al., 2018; Natrass et al., 2020;

Dellinger et al., 2021; Grente, 2021); also (Krofel et al., 2011) and coyotes (*C. latrans*) (Conner et al., 1998). Although this research may be criticized for following less than gold standard designs and for using data collected for other purposes, their results are stronger than correlational, because one or more of their analyses controlled some of the critical variables of location, timing, and the identity of the individuals involved in incidents (Treves et al., 2016, 2019).

We acknowledge that designing robust experiments to determine whether killing carnivores alleviates risks for remaining livestock is difficult, in part because of the rarity and unpredictable timing of such events and the general rapid reactionary management response to them. It is also made more difficult because of the difficulty in differentiating scavenging from predation, the massive areas over which such events occur, the complex interactions of multiple actors including human owners, and the ethical obstacles to performing rigorous experimental studies (e.g., killing carnivores or leaving domestic animals vulnerable to predation). Nevertheless, the most common result of systematic reviews on the outcomes of removal of carnivores on future livestock predation is no effect (e.g. Greentree et al., 2000; Treves et al., 2016; Lennox et al., 2018; Moreira-Arce et al., 2018; van Eeden et al., 2018a; van Eeden et al., 2018b; Khorozyan and Waltert, 2019; Treves et al., 2019; Khorozyan and Waltert, 2020; Lorand et al., 2022; Khorozyan, 2022). Given the challenges of directly studying the effects of carnivore removal on risks for livestock, we believe we can also learn from careful

* Corresponding author at: Panthera, 8 West 40th Street, 18th Floor, New York, NY 10018, USA.

E-mail address: melbroch@panthera.org (L.M. Elbroch).

¹ These authors contributed equally to the paper.

<https://doi.org/10.1016/j.biocon.2023.110106>

Received 5 January 2023; Received in revised form 12 April 2023; Accepted 25 April 2023

Available online 2 May 2023

0006-3207/© 2023 The Author(s). Published by Elsevier Ltd. This is an open access article under the CC BY-NC-ND license (<http://creativecommons.org/licenses/by-nc-nd/4.0/>).

observation of wild carnivore, wild prey, and domestic animal behavior after ‘natural’ experiments involving the removal of predators.

Here, we present an essay to stimulate discussions among ecologists and conservation practitioners about the potential consequences of carnivore removals for livestock owners. We are not arguing against the targeted removals of known repeat offenders, as in some scenarios, this strategy can reduce future threats to people and domestic animals. Instead, our goals are 1) to encourage hypothesis testing of carnivore and prey behaviors which may undermine or counter the effectiveness of interventions intended to protect domestic animals (whether lethal removal or through live-capture and long-term removal from the wild), 2) to encourage a methodological change in the scale of analyses relevant to carnivore interactions with domestic animals (or persons) from populations to smaller social networks, and 3) to encourage public education that includes discussions about the potential consequences of killing carnivores for communities with livestock (e.g. Peebles et al., 2013) and broader appreciation and outreach about the risks of removal, so public policy debate includes both the costs and benefits of living with carnivores (Treves et al., 2009; López-Bao et al., 2017a,b; Gilbert et al., 2021).

Below, we articulate five non-exclusive, likely additive hypotheses for the biological mechanisms that might explain why carnivore removals could fail to affect or even increase risks of future carnivore-livestock conflict (Fig. 1). We focus our hypotheses on the much higher rate of encounters between carnivores and potential prey, rather than the actual rates of injury or death for livestock (Chavez and Gese, 2006; Ohrens et al., 2019). We assume that as encounter rates between carnivores and livestock increase, so too does predation risk for domestic animals. The correlation between encounter rate and attack rate is not perfect of course, however, predator ecologists typically assume that carnivores maximize their encounter rate with their preferred prey, while also sometimes killing alternative prey they encounter while

searching for preferred prey (Schaller, 1972; Mitchell and Lima, 2002; Hayward and Kerley, 2005; Moa et al., 2006; Cristescu et al., 2019). Our hypotheses are not exhaustive, and some hypotheses are better suited to some carnivores over others, but they are all testable and scientifically-derived speculation supported to different degrees by the published literature on carnivores and either domestic or wild prey. Our aim is to stimulate more conversations that align with global concerns over biodiversity loss, and the prevention of livestock attacks over reactive killing of carnivores (Lorand et al., 2022).

Hypothesis 1. Lethal removal may increase local carnivore density and change the age-structure of carnivore social networks. These changes may in turn increase total encounter rates with domestic animals.

Lethal removal creates a vacancy on the landscape, and following species-specific time lags, a greater number of new carnivores may immigrate in to fill the void than the original number of residents that were removed (Adams et al., 2008; Cooley et al., 2009a; Cooley et al., 2009b; Minnie et al., 2016). The allure of vacant habitat may also attract residents of neighboring ranges to shift their territories or expand them. Some carnivores also respond to heavy mortality by increasing litter sizes and reducing the average age of initial female reproduction (e.g., black-backed jackals, Minnie et al., 2016).

As is characteristic of metapopulation dynamics among carnivores, new immigrants are often young animals seeking areas to establish territories (Adams et al., 2008; Cooley et al., 2009a; Cooley et al., 2009b; Minnie et al., 2016). Younger carnivores may exhibit different diets than older animals (African lions, *Panthera leo*, and pumas, Hayward et al., 2007; Elbroch et al., 2017a), which may increase encounter rates with domestic animals. Some evidence also suggests that younger carnivores interact more often with people and domestic animals (Haber, 1996; Linnell et al., 1999; Mattson et al., 2011; Peebles et al., 2013). Increased

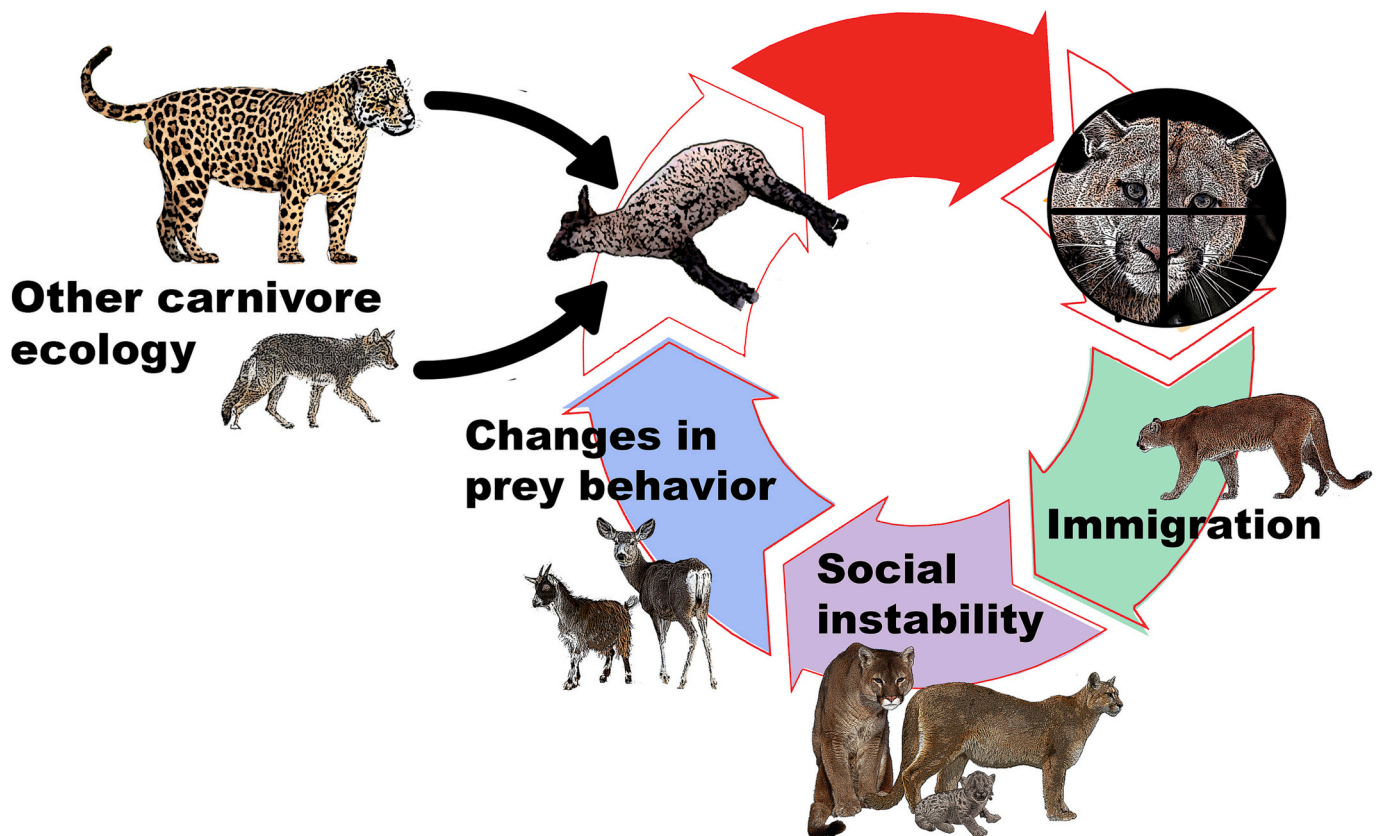


Fig. 1. Conceptual model representing some potentially-additive behavioral factors that might nullify the effects of carnivore removal on mitigating carnivore-livestock conflict, or even exacerbate it.

carnivore density and changes in the age structure of local carnivores may also change intraspecific competition dynamics and social networks, discussed below.

Hypothesis 2. New carnivores, whether immigrants or former neighbors, are unfamiliar with the landscape and local prey distributions. Exploratory movements may increase encounter rates with domestic animals.

Resident carnivores generally prefer wild over domestic prey (Meriggi and Lovari, 1996; Moa et al., 2006; Khorozyan et al., 2015). Many carnivores kill alternative prey, including domestic animals, opportunistically as they encounter them, rather than seek them out (e.g., pumas (Aldredge et al., 2019; Cristescu et al., 2019) and Eurasian lynx (*Lynx lynx*) (Moa et al., 2006; Odden et al., 2008). That does not necessarily imply random encounter rates with domestic animals, as livestock predation is often highly predictable in space and time (Herfindal et al., 2005; Moa et al., 2006; Kaartinen et al., 2009; Kissling et al., 2009; Treves et al., 2011; Davie et al., 2014; Miller et al., 2015; Treves and Rabenhorst, 2017), but it may mean that encounter rates are driven by carnivore behaviors other than hunting. For example, cheetah encounter rates with livestock can be driven by social behaviors (Melzheimer et al., 2020), and dispersal into unknown areas may precipitate carnivore-livestock interactions as well (Linnell et al., 1999).

Carnivores unfamiliar with local prey distributions and activity patterns may search more widely for their food, which may increase encounter rates with alternative prey, including livestock (Fritts et al., 1985; Linnell et al., 1997; Bunnefeld et al., 2006). Carnivores exploring new areas may also spend more time near domestic animals, reducing the availability of their preferred wild prey relative to domestic animals (i.e. the ratio wild:domestic animals), which would also increase encounter rates with domestic versus wild animals (Moa et al., 2006; Khorozyan et al., 2015). Research exploring the effects of native to wild prey ratios on livestock predation have yielded contrary results (e.g. more wild prey reduces conflict Moa et al., 2006; Khorozyan et al., 2015; more wild prey increases conflict via apparent competition, Treves et al., 2004; Suryawanshi et al., 2013). Lower native prey availability may not only lead to greater intraspecific competition, but also greater interspecific dietary overlap and competition (Palacios et al., 2012).

Livestock are also generally easy to locate and more vulnerable to attack than wild prey (Ogada et al., 2003; Wilkinson et al., 2020). Should carnivores experience stress (e.g. hunger or disturbance by people), they may change their foraging behavior, resulting in unfamiliarity with the distributions of their preferred prey, and domestic animals that are often low-cost, high-energetic-reward resources, becoming more attractive. Hungrier pumas, for example, are more likely to risk foraging in suburban areas near people (Blecha et al., 2018).

Hypothesis 3. Lethal removal may destabilize cooperative relationships and social organization among resident carnivores. Social instability may lead to changes in both social behavior and foraging patterns that impact encounter rates with domestic animals.

Social relationships facilitate reproduction, as well as the cooperative behaviors of defending young, hunting larger prey, and defense of territories used by multiple individuals. Even solitary foragers display a variety of social relationships. Felids, for example, exhibit stable, long-lasting social relationships, even while individuals spend the majority of their time alone, e.g. pumas (Elbroch et al., 2017b); jaguars, *Panthera onca* (Jędrzejewski et al., 2022), leopards, *Panthera pardus* (Bailey, 1993); domestic cats, *Felis catus* (MacDonald et al., 1987). Here, we use social organization to mean the full range of possible relationships and affiliative bonds that influence intraspecific avoidance and aggressive interactions, with the assumption that these are influenced by individual cognition, personalities and cultures within families and other social networks (Hare and Tomasello, 2005; MacLean and Hare, 2015; Marshall-Pescini et al., 2017). We define instability in social organization as a disruption of existing relationships necessitating

reorganization, assessment, formation of new relationships, and possible aggression.

Some social changes may not directly influence encounter rates with livestock, but instead cause stress that could indirectly impact diverse behaviors. For example, when resident female carnivores lose a resident male to lethal removal, they may face new risks of infanticide following the immigration of new males (Pusey and Packer, 1993; Swenson et al., 1997; Packer et al., 2009), or they may experience lost mating opportunities. Dependent young that lose a parent to lethal removal may lose associated opportunities to learn appropriate prey and prey handling (Caro, 1987; Caro and Hauser, 1992; Elbroch and Quigley, 2013), making livestock potentially more attractive (Wilkinson et al., 2020, though refuted in Linnell et al., 1999). A group of social carnivores that suffers the loss of a cooperator to lethal removal may experience escalated competition and intraspecific killing by rival groups for territory, food patches, and mates (Gittleman, 1989; Packer et al., 1990). As a result, small coalitions may find themselves displaced or injured by larger coalitions within either sex (Manson and Wrangham, 1991; McComb et al., 1994). Stress and injuries among carnivores in an unstable social network may make domestic animals more attractive, although years of research report equivocal evidence for this prediction (Linnell et al., 1999).

Changes in a carnivore's social environment might necessitate changes in a suite of behaviors for the remaining conspecifics, depending on an individual's life history stage, competitive ability, and familiarity with the newly vacant habitat (Swenson et al., 1997; Minnie et al., 2016). These and other changes may affect individual ranging behavior, prey selection, and therefore encounter rates with domestic animals. For example, large wolf packs that experience removals and that split into several smaller packs may kill more prey in terms of biomass than the original pack (Zimmermann et al., 2015). In another study, wolf pack size was negatively associated with the frequency of attacks on livestock and positively associated with aggressive encounters with hounds (Wydeven et al., 2004).

Hypothesis 4. Reductions in the abundance and distribution of one carnivore species may change the abundance or distribution of another carnivore species, resulting in increased encounter rates with domestic prey.

Carnivore species rarely exist in isolation, such that killing a dominant carnivore may lead subordinate species to prey on domestic animals more than did the carnivore that was removed. This perverse outcome was suggested anecdotally 65 years ago (Newby and Brown, 1958; Natrass et al., 2020). Top carnivores sometimes constrain mesopredator distributions (Levi and Wilmers, 2012; Newsome et al., 2017; Ruprecht et al., 2021), which in turn constrain smaller carnivores (Levi and Wilmers, 2012); probably all carnivores sometimes kill domestic animals. When large carnivores are eliminated, some evidence suggests that smaller carnivores benefit (Prugh et al., 2009; Levi and Wilmers, 2012), however, conclusive scientific evidence for mesopredator release can be elusive at local scales (Crooks and Soulé, 1999; Krofel et al., 2007; Allen et al., 2016; Crimmins and Van Deelen, 2019). In systems with multiple large carnivores, even apex carnivores are affected by the presence of other, more-dominant apex species (e.g., leopards by tigers, Harihar et al., 2011; pumas by wolves, Elbroch and Kusler, 2017; Elbroch et al., 2020; coyotes by pumas and wolves, Levi and Wilmers, 2012; Ruprecht et al., 2021). Any changes in the local abundance and distributions of one carnivore, especially if maintained over time by management, will impact the distributions and abundances of other carnivores, and these shifting community dynamics will influence total carnivore-livestock encounter rates.

Hypothesis 5. Lethal removal may precipitate changes in native prey or domestic animal behavior that makes livestock more vulnerable to predation.

Risk allocation theory (Lima and Dill, 1990; Lima and Bednekoff,

1999) assumes that prey species can detect the presence or absence of resident carnivores, and that they respond by changing their own distributions and behaviors to minimize the chance of encountering them. When carnivores are removed, we assume that wild and domestic prey may respond by changing their distributions and other behaviors, and that these changes will also impact their encounter rates with remaining carnivores. This hypothesis best applies to free range livestock on public lands, and private ranches or other large enclosures, where they overlap more with wild prey and their movements are not entirely dictated by people. Low-stress livestock-handling LSLH (sensu Louchouart and Treves, 2023), for example, enhances innate livestock behavior to protect livestock. Livestock owners and managers detect carnivore sign and signs of anxiety in cattle and sheep, and then move their livestock into safer aggregations or safer areas.

2. Conclusions

While wildlife managers sometimes kill carnivores to mitigate the social consequences of carnivore-livestock conflict (Linnell, 2011; Swan et al., 2017), the assumption that killing carnivores reduces livestock predation is no longer tenable in general terms (Lorand et al., 2022). Under some circumstances, there may even be unexpected consequences for lethally removing carnivores (Woodroffe et al., 2006; Peebles et al., 2013; Natrass et al., 2020). The counterargument is that managers need to apply greater pressure to reduce carnivore populations enough to achieve measurable outcomes, but this runs counter to findings showing no effect of heavy, long-term carnivore control to increase wild ungulate abundance (Hurley et al., 2011; Clark and Hebblewhite, 2021; Miller et al., 2022). Prey selection, predator-prey dynamics, and human-wildlife interactions are complex issues with multiple causal factors difficult to tease apart. For example, if many conditions predicted in our above five hypotheses are met, we would expect additive or multiplicative effects raising encounter rates with domestic animals (Fig. 1). This implies that there can be consequences for people when we remove carnivores, and there are certainly consequences for individual carnivores, and their social networks and populations. Human-induced mortality very likely has the same or similar effects on carnivore social organization as other causes of death, but is much more frequent (Woodroffe and Ginsberg, 1998; Wydeven et al., 2001; Treves et al., 2017). Repeated lethal removal of carnivores may also result in more than additive effects on carnivore social organization and behavior. For example, male pumas exhibit greater home range overlap in heavily hunted populations (Maletzke et al., 2014), which may increase opportunities for intraspecific aggression that further destabilizes carnivore social networks, jackals and other carnivores respond to unstable territoriality and social networks with increased fecundity (Minnie et al., 2016), and culling European badgers (*Meles meles*) disrupts their social networks, increases their social interactions, and increases disease transmission within their populations (Woodroffe et al., 2006).

The half-century debate about the effects of carnivore-removal on livestock conflict examined at a population-scale highlights the importance of approaching the question via well-designed experimental approaches, and at the appropriate scale. Individual carnivores or social groups prey on livestock. Therefore, the social network of carnivores is the appropriate scale of analysis for evaluating the effect of interventions to protect domestic animals. For example, dissatisfaction with the equivocal conclusions drawn from population-scale, correlational analyses of wolf removal (Wielgus and Peebles, 2014; Poudyal et al., 2016; Kompaniyets and Evans, 2017) has led to a gradual acceptance of the need for smaller-scale analyses that compare wolf social units or individual territories (Bradley et al., 2015; Santiago-Avila et al., 2018; Grente, 2021). In turn, these are giving way to the even more precise randomized, controlled trials that focus on replicated social units (i.e. herds of livestock, packs of wolves; Santiago-Avila et al., 2018; Louchouart and Treves, 2023). The field is increasingly turning to more robust designs, either through random-assignment to controls and

treatments, before-after-control-impact designs, or through exhaustive statistical control over autocorrelation and potentially confounding variables (Treves et al., 2019; Khorozyan, 2022). Further, the tools available today to test our hypotheses and others are more powerful than ever before. GPS technology and motion-triggered cameras can measure the frequency with which carnivores enter pastures or enclosures, as well as their ranging behavior, territoriality and more (Kays et al., 2015). Proximity sensors and camera collars worn by carnivores and wild prey and livestock can measure encounter rates among and within species (Prugh et al., 2019), and genetic methods provide us tools to determine the carnivore that attacked or killed livestock (Williams et al., 2003), as might be useful in testing whether removing one carnivore species sees an increase in predation by another. Such research could be done strategically where carnivore manipulation is already underway, as is the case in many western states in the USA.

Killing a carnivore should not be attempted without first considering the potential costs and benefits for domestic animals, people, and sympatric wildlife, including surviving conspecifics. In general, wildlife managers present lethal removal to society as a means of achieving a measurable reduction in the risks posed to livestock (Linnell, 2011). In other words, lethal removal is presented as a best strategy to aid individual livestock owners (e.g. Conover, 2001), which especially makes sense within the context of global north societies, many of which prioritize individualistic world views (Gerlach et al., 2018). There are however, potential negative consequences for livestock owners, as presented above, and definite consequences for carnivores and their ecosystem functions contributing to biodiversity and ecological resilience (Enquist et al., 2020; LaBarge et al., 2022). Therefore, lethal removal has negative impacts on communities, both through potential carnivore social disruption that may impact the safety of other people's livestock, as well as the cascading impacts of carnivore removal on ecosystem health (Woodroffe et al., 2006; Estes et al., 2011; Enquist et al., 2020).

We strongly encourage wildlife managers to present and openly discuss the potential costs alongside the potential benefits of lethal removal to both individual livestock owners and their larger communities, which better sets the stage for democratic conservation decision making inclusive of the diverse views and societal goals reflective of the world's cultures (Treves and Santiago-Avila, 2020; Lele, 2021). Even when managers decide to remove a carnivore, the identity of the offending animal(s) should be ascertained with great confidence lest a livestock killer survive and a non-culprit be killed (sensu Graham et al., 2011; López-Bao et al., 2017a,b). One way to address the uncertainties we summarized here is for authorities to monitor the after-effects of killing a carnivore on carnivore social networks (e.g., immigrants, social instability), as well as prey distributions and behaviors, just as one measures the effects of an experimental manipulation on all subjects. Reporting the effects of lethal interventions to the public and domestic animal owners should be an essential step in such actions, as should the methods used for evaluation. Results from robust research designs should be given more emphasis than data derived from simple correlations or before-after designs lacking the addition of control-impact (Khorozyan, 2022).

Given the tremendous uncertainty about killing wild carnivores, a prudent choice may be to preferentially select preventive, non-lethal methods of protecting livestock proven by randomized, controlled trials (RCTs) (Lorand et al., 2022). Non-lethal methods have been found effective in protecting livestock from wild carnivores in numerous situations (e.g., painted eye-spots on cattle versus African lions (Radford et al., 2020); low-stress livestock handling by 'range riders' to protect cattle from pumas, grizzly bears *U. arctos*, black bears, gray wolves, and coyotes (Louchouart and Treves, 2023); fladry flagging against wolves and coyotes (Davidson-Nelson and Gehring, 2010; Young et al., 2019).

Currently non-lethal methods have been tested with higher standards than have lethal (Treves et al., 2016; van Eeden et al., 2018a; Treves et al., 2019; Khorozyan and Waltert, 2020; Khorozyan, 2022).

Combining livestock defenses has also been advocated for decades (Linhart, 1981; Shivik, 2006; Stone et al., 2017; Fergus, 2020). One might argue that lethal methods should be paired with one or more protective husbandry methods, so that if the killing is not effective the husbandry may succeed. This argument begs an experimental test, because lethal methods should face the same burden of proof as non-lethal interventions.

Culling carnivores and lethal removals remain the leading strategy to address carnivore-livestock conflict even while accumulated science suggests that non-lethal approaches are more effective (Lorand et al., 2022). We believe this is in part due to the lag time between scientific discovery and the dissemination of new information to policy makers and the general public (Messmer et al., 2001). Therefore, simultaneous with new hypothesis testing to determine when and why the lethal control of carnivores may increase rather than reduce risks for livestock, we call on wildlife managers and conservation practitioners to ensure their outreach about conflict includes the potential consequences of lethal carnivore control for local livestock, and both the costs and benefits of living sympatric with carnivores and intact ecosystems.

Declaration of competing interest

The authors declare no competing interests.

Data availability

No data was used for the research described in the article.

Acknowledgements

This work did not receive any specific grant from funding agencies in the public, commercial, or not-for-profit sectors.

References

- Adams, L.G., Stephenson, R.O., Dale, B.W., Ahgook, R.T., Demma, D.J., 2008. Population dynamics and harvest characteristics of wolves in the Central Brooks Range, Alaska. *Wildl. Monogr.* 170, 1–25.
- Allredge, M.W., Buderman, F.E., Blecha, K.A., 2019. Human–cougar interactions in the wildland–urban interface of Colorado’s front range. *Ecol. Evol.* 9, 1–17.
- Allen, B.L., Lundie-Jenkins, G., Burrows, N.D., Engeman, R.M., Fleming, P.J.S., Leung, L. K.P., 2016. Does lethal control of top-predators release mesopredators? A re-evaluation of three Australian case studies. *Ecol. Manag. Restor.* 15, 193–195.
- Bailey, T.N., 1993. *The African Leopard: Ecology and Behavior of a Solitary Felid*. Columbia University Press, New York.
- Blecha, K.A., Boone, R.B., Allredge, M.W., 2018. Hunger mediates apex predator’s risk avoidance. *J. Anim. Ecol.* 87, 609–622.
- Bradley, E.H., Robinson, H.S., Bangs, E.E., Kunkel, K.E., Jimenez, M.D., Gude, J.D., Grimm, T., 2015. Effects of wolf removal on livestock depredation recurrence and wolf recovery in Montana, Idaho, and Wyoming. *J. Wildl. Manag.* 79, 1337–1346.
- Bunnefeld, N., Linnell, J.D.C., Odden, J., van Duijn, M.A.J., et al., 2006. Risk taking by eurasian lynx (*Lynx lynx*) in a human-dominated landscape: effects of sex and reproductive status. *J. Zool.* 270, 31–39.
- Caro, T.M., 1987. Indirect costs of play: cheetah cubs reduce maternal hunting success. *Anim. Behav.* 35, 295–297.
- Caro, T.M., Hauser, M.D., 1992. Is there teaching in nonhuman animals? *Q. Rev. Biol.* 67, 151–174.
- Chavez, A.S., Gese, E.M., 2006. Landscape use and movements of wolves in relation to livestock in a wildland–agriculture matrix. *J. Wildl. Manag.* 70, 1079–1086.
- Clark, T.J., Hebblewhite, M., 2021. Predator control may not increase ungulate populations in the future: a formal meta-analysis. *J. Appl. Ecol.* 58, 812–824.
- Conner, M.M., Jaeger, M.M., Weller, T.J., McCullough, D.R., 1998. Effect of coyote removal on sheep depredation in northern California. *J. Wildl. Manag.* 62, 690–699.
- Conover, M.R., 2001. Effect of hunting and trapping on wildlife damage. *Wildl. Soc. Bull.* 29, 521–532.
- Cooley, H.S., Wielgus, R.B., Koehler, G.M., Maletzke, B.T., 2009a. Source populations in carnivore management: cougar demography and emigration in a lightly hunted population. *Anim. Conserv.* 12, 321–328.
- Cooley, H.S., Wielgus, R.B., Robinson, H.S., Koehler, G.M., Maletzke, B.T., 2009b. Does hunting regulate cougar populations? A test of the compensatory mortality hypothesis. *Ecology* 90, 2913–2921.
- Crimmins, S., Van Deelen, T.R., 2019. Limited evidence for mesocarnivore release following wolf recovery in WisconsinUSA. *Wildl. Biol.* 1, 1–7.
- Cristescu, B., Bose, S., Elbroch, L.M., Allen, M.L., Wittmer, H.U., 2019. Habitat selection when killing primary versus alternative prey species supports prey specialization in an apex predator. *J. Zool.* 309, 259–268.
- Crooks, K.R., Soulé, M.E., 1999. Mesopredator release and avifaunal extinctions in a fragmented system. *Nature* 400, 563–566.
- Davidson-Nelson, S.J., Gehring, T.M., 2010. Testing fidelity as a nonlethal management tool for wolves and coyotes in Michigan. *Hum. Wildl. Interact.* 4, 87–94.
- Davie, H.S., Murdoch, J.D., Lhagvasuren, A., Reading, R.P., 2014. Measuring and mapping the influence of landscape factors on livestock predation by wolves in Mongolia. *J. Arid Environ.* 103, 85–91.
- Dellinger, J.A., Macon, D.K., Rudd, J.L., Clifford, D.L., Torres, S.G., 2021. Temporal trend and drivers of mountain lion depredation in California. *Hum. Wildl. Interact.* 15, 162–177.
- The human dimension in addressing conflict with large carnivores. In: Dickman, A., Marchini, S., Manfredo, M., Macdonald, D., Willis, K.J. (Eds.), 2013. *Key Topics in Conservation Biology*, 2. John Wiley & Sons, London.
- van Eeden, L.M., Crowther, M.S., Dickman, C.R., Macdonald, D.W., Ripple, W.J., Ritchie, E.G., Newsome, T.M., 2018a. Managing conflict between large carnivores and livestock. *Conserv. Biol.* 32, 26–34. <https://doi.org/10.1111/cobi.12959>.
- van Eeden, L.M., Eklund, A., Miller, J.R.B., López-Bao, J.V., Cejtin, M.R., Chapron, G., Crowther, M.S., Dickman, C.R., Frank, J., Krolf, M., Macdonald, D.W., McManus, J., Meyer, T.K., Middleton, A.D., Newsome, T.M., Ripple, W.J., Ritchie, E.G., Schmitz, O.J., Stoner, K.J., Tourani, M., Treves, A., 2018b. Carnivore conservation needs evidence-based livestock protection. *PLoS Biol.* 16.
- Elbroch, L.M., Kusler, A.L., 2017. Are pumas subordinate carnivores, and does it matter? *PeerJ* 6, e4293.
- Elbroch, L.M., Quigley, H.B., 2013. Observations of wild cougar kittens with live prey: implications for learning and survival. *Can. Field Nat.* 26, 333–335.
- Elbroch, L., Feltner, J., Quigley, H., 2017a. Stage-dependent puma predation on dangerous prey. *J. Zool.* 302, 164–170.
- Elbroch, L.M., Levy, M.A., Lubell, M., Quigley, H.B., Caragiulo, A., 2017b. Adaptive social behaviors in a solitary carnivore. *Sci. Adv.* 3, e1701218.
- Elbroch, L.M., Fergus, J., Quigley, H., Craighead, D., Thompson, D., Wittmer, H.U., 2020. Reintroduced wolves and hunting limit the abundance of a subordinate apex predator in a multi-use landscape. *Proc. R. Soc. B* 287, 20202202.
- Enquist, B.J., Abraham, A.J., Harfoot, M.B., Malhi, Y., Doughty, C.E., 2020. The megabiota are disproportionately important for biosphere functioning. *Nat. Comm.* 11, 699.
- Estes, J.A., Terborgh, J., Brashares, J.S., Power, M.E., Berger, J., Bond, W.J., Carpenter, S.R., Essington, T.E., Holt, R.D., Jackson, J.B., Marquis, R.J., Oksanen, L., Oksanen, T., Paine, R.T., Pikitch, E.K., Ripple, W.J., Sandin, S.A., Scheffer, M., Schoener, T.W., Shurin, J.B., Sinclair, A.R., Soulé, M.E., Virtanen, R., Wardle, D.A., 2011. Trophic downgrading of planet earth. *Science* 333, 301–306.
- Fergus, A.R., 2020. Building Carnivore Coexistence on Anishinaabe Land: Gold Standard Non-lethal Deterrent Research and Relationship Building Between Livestock Farmers and the Bad River Band of the Lake Superior Tribe of Chippewa Indians. University of Wisconsin, Madison, WI.
- Fritts, S.H., Paul, W.J., Mech, L.D., 1985. Can relocated wolves survive? *Wildl. Soc. Bull.* 13, 459–463.
- Gerlach, A.J., Teachman, G., Laliberte-Rudman, D., Aldrich, R.M., Huot, S., 2018. Expanding beyond individualism: engaging critical perspectives on occupation. *Scand. J. Occup. Ther.* 25, 35–43.
- Gilbert, S., Carter, N., Naidoo, R., 2021. Predation services: quantifying societal effects of predators and their prey. *Front. Ecol. Environ.* 19, 292–299.
- Gittleman, J.L., 1989. Carnivore group living: comparative trends. In: Gittleman, J.L. (Ed.), *Carnivore Behavior, Ecology and Evolution*, vol. 1. Comstock Assoc, Ithaca, pp. 183–207.
- Graham, I.M., Harris, R.N., Matejusova, I., Middlemas, S.J., 2011. Do ‘rogue’ seals exist? Implications for seal conservation in the UK. *Anim. Conserv.* 14, 587–598.
- Greentree, C., Saunders, G., McLeod, L., Hone, J., 2000. Lamb predation and fox control in South-Eastern Australia. *J. Appl. Ecol.* 37, 935–943.
- Grete, O., 2021. *résentation des objectifs et de la méthodologie de la thèse sur l’efficacité des tirs de loup et la gestion adaptative du loup, menée conjointement par l’ONCFS et le CEFE*. Gières, France.
- Haber, G.C., 1996. Biological, conservation, and ethical implications of exploiting and controlling wolves. *Conserv. Biol.* 10, 1068–1081.
- Hare, B., Tomasello, M., 2005. Human-like social skills in dogs? *Trends Cogn. Sci.* 9, 439–444.
- Harihar, A., Pandav, B., Goyal, S.P., 2011. Responses of leopard *Panthera pardus* to the recovery of a tiger *Panthera tigris* population. *J. Appl. Ecol.* 48, 806–814.
- Hayward, M.W., Kerley, G.I.H., 2005. Prey preferences of the lion (*Panthera leo*). *J. Zool.* 207, 309–322.
- Hayward, M.W., Hofmeyr, M., O’Brien, J., Kerley, G.I.H., 2007. Testing predictions of the prey of the lion (*Panthera leo*) derived from modelled prey preferences. *J. Wildl. Manag.* 71, 1567–1575.
- Herfindal, I., Linnell, J.D.C., Moa, P.F., Odden, J., Austmo, L.B., Andersen, R., 2005. Does recreational hunting of lynx reduce depredation losses of domestic sheep? *J. Wildl. Manag.* 69, 1034–1042.
- Hurley, M.A., Unsworth, J.W., Zager, P., Hebblewhite, M., Garton, E.O., Montgomery, D. M., Skalski, J.R., Maycock, C.L., 2011. Demographic response of mule deer to experimental reduction of coyotes and mountain lions in southeastern Idaho. *Wildl. Monogr.* 178, 1–33.
- Jędrzejewski, W., Hoogesteijn, R., Devlin, A.L., et al., 2022. Collaborative behaviour and coalitions in male jaguars (*Panthera onca*)—evidence and comparison with other felids. *Behav. Ecol. Sociobiol.* 76, 121.

- Kaartinen, S., Luoto, M., Kojola, I., 2009. Carnivore-livestock conflicts: determinants of wolf (*Canis lupus*) depredation on sheep farms in Finland. *Biodivers. Conserv.* 18, 3503–3517.
- Kays, R., Crofoot, M.C., Jetz, W., Wikelski, M., 2015. Terrestrial animal tracking as an eye on life and planet. *Science* 348, aza2478.
- Khorozyan, I., 2022. Defining practical and robust study designs for interventions at terrestrial mammalian predators. *Conserv. Biol.* 36, e13805.
- Khorozyan, I., Waltert, M., 2019. How long do anti-predator interventions remain effective? Patterns, thresholds and uncertainty. *R. Soc. Open Sci.* 6.
- Khorozyan, I., Waltert, M., 2020. Variation and conservation implications of the effectiveness of anti-bear interventions. *Sci. Rep.* 10, 15341.
- Khorozyan, I., Ghoddousi, A., Soofi, M., Waltert, M., 2015. Big cats kill more livestock when wild prey reaches a minimum threshold. *Biol. Conserv.* 1925, 268–275.
- Kissling, W.D., Fernandez, N., Paruelo, J.M., 2009. Spatial risk assessment of livestock exposure to pumas in Patagonia, Argentina. *Ecography* 32, 807–817.
- Kompaniyets, L., Evans, M., 2017. Modeling the relationship between wolf control and cattle depredation. *PLoS ONE* 12, e0187264.
- Krofel, M., Giannatos, G., Ćirović, D., Stoyanov, S., Newsome, T.M., 2007. Golden jackal expansion in Europe: a case of mesopredator release triggered by continent-wide wolf persecution? *Hystrix* 28, 9–15.
- Krofel, M., Černe, R., Jerina, K., 2011. Effectiveness of wolf (*Canis lupus*) culling as a measure to reduce livestock depredations. *Acta Silv. Ligni* 95, 11–22.
- LaBarge, L., Evans, M.J., Miller, J.R.B., Cannataro, G., Hunt, C., Flemming, E., Elbroch, L.M., 2022. Pumas as ecological brokers: a review of their biotic relationships. *Mamm. Rev.* 52, 360–376.
- Lele, S., 2021. From wildlife-ism to ecosystem-service-ism to a broader environmentalism. *Environ. Conserv.* 48, 5–7.
- Lennox, R.J., Gallagher, A.J., Ritchie, E.G., Cooke, S.J., 2018. Evaluating the efficacy of predator removal in a conflict-prone world. *Biol. Conserv.* 224, 277–289.
- Levi, T., Wilmers, C.C., 2012. Wolves-coyotes-foxes: a cascade among carnivores. *Ecology* 93, 921–929.
- Lima, S.L., Bednekoff, P.A., 1999. Temporal variation in danger drives antipredator behavior: the predation risk allocation hypothesis. *Am. Nat.* 153, 649–659.
- Lima, S.L., Dill, L.M., 1990. Behavioral decisions made under the risk of predation: a review and prospectus. *Can. J. Zool.* 68, 619–640.
- Linhart, S.B., 1981. Managing coyote damage problems with non-lethal techniques: recent advances in research. In: *Proceedings of the Eastern Wildlife Damage Control Conference*, 1, pp. 105–118.
- Linnell, J.D.C., 2011. Can we separate the sinners from the scapegoats? *Anim. Conserv.* 14, 602–603.
- Linnell, J.D.C., Aanes, R., Swenson, J.E., Odden, J., Smith, M.E., 1997. Translocation of carnivores as a method for managing problem animals: a review. *Biodivers. Conserv.* 6, 1245–1257.
- Linnell, J.D.C., Odden, J., Smith, M.E., Aanes, R., Swenson, J.E., 1999. Large carnivores that kill livestock: do problem individuals really exist? *Wildl. Soc. Bull.* 27, 698–705.
- López-Bao, J.V., Bruskotter, J., Chapron, G., 2017a. Finding space for large carnivores. *Nat. Ecol. Evol.* 1, 1–2.
- López-Bao, J.V., Frank, J., Svensson, L., Åkesson, M., Langefors, Å., 2017b. Building public trust in compensation programs through accuracy assessments of damage verification protocols. *Biol. Conserv.* 213, 36–41.
- Lorand, C., Robert, A., Gastineau, A., Mihoub, J.B., Bessa-Gomes, C., 2022. Effectiveness of interventions for managing human-large carnivore conflicts worldwide: scare them off, don't remove them. *Sci. Total Environ.* 838, 156195.
- Louchouart, N.X., Treves, A., 2023. Low-stress livestock handling protects cattle in a five-predator habitat. *PeerJ* 11, e14788.
- MacDonald, D.W., Apps, P.J., Carr, G.M., Kerby, G., 1987. Social dynamics, nursing coalitions and infanticide among farm cats, *Felis catus*. In: Burghardt, W.P.G.M., Wickler, W. (Eds.), *Advanced Ethology*. Verlag, Berlin.
- MacLean, E.L., Hare, B., 2015. Dogs hijack the human bonding pathway. *Science* 348, 280–281.
- Maletzke, B.T., Wielgus, R., Koehler, G.M., Swanson, M., Cooley, H., Allredge, J.R., 2014. Effects of hunting on cougar spatial organization. *Ecol. Evol.* 4, 2178–2185.
- Manson, J.H., Wrangham, R.W., 1991. Intergroup aggression in chimpanzees and humans. *Curr. Anthropol.* 32, 369–390.
- Marshall-Pescini, S., Schwarz, J.F.L., Kostelnika, I., Virányi, Z., Range, R., 2017. Importance of a species' socioecology: wolves outperform dogs in a conspecific cooperation task. *Proc. Acad. Nat. Sci.* 114, 11793–11798.
- Mattson, D., Logan, K., Sweanor, L., 2011. Factors governing risk of cougar attacks on humans. *Hum. Wildl. Interact.* 5, 135–158.
- McComb, K., Packer, C., Pusey, A.E., 1994. Roaring and numerical assessment in contests between groups of female lions, *Panthera leo*. *Anim. Behav.* 47, 379–387.
- Melzheimer, J., Heinrich, S.K., Wasiolka, B., Mueller, R., Thalwitzer, S., 2020. Communication hubs of an asocial cat are the source of a human – carnivore conflict and key to its solution. *Proc. Acad. Nat. Sci.* 117, 3325–3333.
- Meriggi, A., Lovari, S., 1996. A review of wolf predation in southern Europe: does the wolf prefer wild prey to livestock? *J. Appl. Ecol.* 33, 1561–1571.
- Messmer, T.A., Reiter, D., West, B.C., 2001. Enhancing wildlife sciences' linkage to public policy: lessons from the predator-control pendulum. *Wildl. Soc. Bull.* 29, 1253–1259.
- Miller, J.R.B., Jhala, Y.V., Jena, J., Schmitz, O.J., 2015. Landscape-scale accessibility of livestock to tigers: implications of spatial grain for modeling predation risk to mitigate human–carnivore conflict. *Ecol. Evol.* 5, 1354–1367.
- Miller, S.D., Person, D.K., Bowyer, R.T., 2022. Efficacy of killing large carnivores to enhance moose harvests: new insights from a long-term view. *Diversity* 14, 939.
- Minnie, L., Gaylard, A., Kerley, G.I.H., 2016. Compensatory life-history responses of a mesopredator may undermine carnivore management efforts. *J. Anim. Ecol.* 53, 379–387.
- Mitchell, W.A., Lima, S.L., 2002. Predator-prey shell games: large-scale movement and its implications for decision-making by prey. *Oikos* 99, 249–259.
- Moa, P.F., Herfindal, I., Linnell, J.D.C., Overskaug, K., Kvam, T., Andersen, R., 2006. Does the spatiotemporal distribution of livestock influence forage patch selection in Eurasian lynx *Lynx lynx*? *Wildl. Biol.* 2, 63–70.
- Moreira-Arce, D., Ugarte, C.S., Zorondo-Rodríguez, F., Simonetti, J.A., 2018. Management tools to reduce carnivore-livestock conflicts: current gap and future challenges. *Rangel. Ecol. Manag.* 71, 389–394.
- Natrass, N., Conradie, B., Stephens, J., Drouilly, M., 2020. Culling recolonizing mesopredators increases livestock losses: evidence from the South African Karoo. *Ambio* 49, 1222–1231.
- Naughton-Treves, L., Grossberg, R., Treves, A., 2003. Paying for tolerance: the impact of livestock depredation and compensation payments on rural citizens' attitudes toward wolves. *Conserv. Biol.* 17, 1500–1511.
- Newby, F., Brown, R., 1958. A new approach to predator management in Montana. *Montana Wildl.* 8, 22–27.
- Newsome, T.M., Greenville, A.C., Ćirović, D., Dickman, C.R., Johnson, C.N., Krofel, M., Letnic, M., Ripple, W.J., Ritchie, E.G., Stoyanov, S., Wirsing, A.J., 2017. Top predators constrain mesopredator distributions. *Nature Comm.* 8, 15469.
- Odden, J., Herfindal, I., Linnell, J.D.C., Andersen, R., 2008. Vulnerability of domestic sheep to lynx depredation in relation to roe deer density. *J. Wildl. Manag.* 72, 276–282.
- Ogada, M.O., Woodroffe, R., Oguge, N.O., Frank, L.G., 2003. Limiting depredation by african carnivores: the role of livestock husbandry. *Conserv. Biol.* 17, 1521–1530.
- Ohrnes, O., Bonacic, C., Treves, A., 2019. Non-lethal defense of livestock against predators: flashing lights deter puma attacks in Chile. *Front. Ecol. Environ.* 17, 32–38.
- Oregon Department of Fish and Wildlife, n.d. Oregon Department of Fish and Wildlife (n.d.), "Hunting Cougar in Oregon," <https://myodfw.com/articles/hunting-cougar-oregon>, accessed August 12, 2019.
- Packer, C., Scheel, D., Pusey, A.E., 1990. Why lions form groups: food is not enough. *Am. Nat.* 136, 1–19.
- Packer, C., Kosmala, M., Cooley, H.S., Brink, H., Pintea, L., Garshelis, D., Purchase, G., Strauss, M., Swanson, A., Balme, G., Hunter, L., Nowell, K., 2009. Sport hunting, predator control and conservation of large carnivores. *PLoS ONE* 4, e5941. <https://doi.org/10.1371/journal.pone.0005941>, 59.
- Palacios, R., Walker, R.S., Novaro, A.J., 2012. Differences in diet and trophic interactions of patagonian carnivores between areas with mostly native or exotic prey. *Mamm. Biol.* 77, 183–189.
- Peebles, K., Wielgus, R.B., Maletzke, B.T., Swanson, M.E., 2013. Effects of remedial sport hunting on cougar complaints and livestock depredations. *PLoS ONE* 8, e79713.
- Poudyal, N., Baral, N., Asah, S.T., 2016. Wolf lethal control and depredations: counter-evidence from respecified models. *PLoS ONE* 11, e0148743.
- Prugh, L.R., Stoner, C.J., Epps, C.W., Bean, W.T., Ripple, W.J., Laliberte, A.S., Brashares, J.S., 2009. The rise of the mesopredator. *Bioscience* 59, 779–791.
- Prugh, L.R., Sivy, K.J., Mahoney, P.J., Ganz, T.R., Dittmer, M.A., van de Kerk, M., Gilbert, S.L., Montgomery, R.A., 2019. Designing studies of predation risk for improved inference in carnivore-ungulate systems. *Biol. Conserv.* 232, 194–207.
- Pusey, A.E., Packer, C., 1993. Infanticide in lions: consequences and counter-strategies. In: Parmigiani, S., von Saal, F.S. (Eds.), *Infanticide and Parental Care*. Harwood Academic Press, London.
- Radford, C.G., McNutt, J.W., Rogers, T., Maslen, B., Jordan, N.R., 2020. Artificial eyespots on cattle reduce predation by large carnivores. *Commun. Biol.* 3, 430.
- Ruprecht, J., Eriksson, C.E., Forrester, T.D., Spitz, D.B., Clark, D.A., Wisdom, M.J., et al., 2021. Variable strategies to solve risk-reward tradeoffs in carnivore communities. *Proc. Acad. Nat. Sci.* 118, e2101614118.
- Santiago-Avila, F.J., Cormman, A.M., Treves, A., 2018. Killing wolves to prevent predation on livestock may protect one farm but harm neighbors. *PLoS ONE* 13, e0189729.
- Schaller, G.B., 1972. *The Serengeti Lion: A Study of Predator-Prey Relations*. University of Chicago Press, Chicago.
- Shivik, J.A., 2006. Tools for the edge: what's new for conserving carnivores. *Bioscience* 56, 253–259.
- Stone, S.A., Breck, S.W., Timberlake, J., Haswell, P.M., Najera, F., Bean, B.S., Thornhill, D.J., 2017. Adaptive use of nonlethal strategies for minimizing wolf-sheep conflict in Idaho. *J. Mammal.* 98, 33–44.
- Suryawanshi, K.R., Bhatnagar, Y.V., Redpath, S., Mishra, C., 2013. People, predators and perceptions: patterns of livestock depredation by snow leopards and wolves. *J. Appl. Ecol.* 50, 550–560.
- Swan, G., Redpath, S.M., Bearhop, S., McDonald, R.A., 2017. Ecology of problem individuals and the efficacy of selective wildlife management. *Trends Ecol. Evol.* 32, 518–530.
- Swenson, J.E., Sandegren, F., Soderberg, A., Bjarvall, A., Franzen, R., Wabakken, P., 1997. Infanticide caused by hunting of male bears. *Nature* 386, 450–451.
- Teichman, K.J., Cristescu, B., Darimont, C.T., 2016. Hunting as a management tool? Cougar-human conflict is positively related to trophy hunting. *BMC Ecol.* 16, 44. <https://doi.org/10.1186/s12898-016-10098-12894>.
- Treves, A., Karanth, K.U., 2003. Human-carnivore conflict and perspectives on carnivore management worldwide. *Conserv. Biol.* 17, 1491–1499.
- Treves, A., Rabenhorst, M.F., 2017. Risk map for wolf threats to livestock still predictive 5 years after construction. *PLoS ONE* 12, e0180043.
- Treves, A., Santiago-Avila, F.J., 2020. Myths and assumptions about human-wildlife conflict and coexistence. *Conserv. Biol.* 34, 811–818.

- Treves, A., Naughton-Treves, L., Harper, E.L., Mladenoff, D.J., et al., 2004. Predicting human-carnivore conflict: a spatial model based on 25 years of wolf predation on livestock. *Conserv. Biol.* 18, 114–125.
- Treves, A., Wallace, R.B., White, S., 2009. Participatory planning of interventions to mitigate human-wildlife conflicts. *Conserv. Biol.* 23, 1577–1587.
- Treves, A., Martin, K.A., Wydeven, A.P., Wiedenhoef, J.E., 2011. Forecasting environmental hazards and the application of risk maps to predator attacks on livestock. *Bioscience* 61, 451–458.
- Treves, A., Krofel, M., McManus, J., 2016. Predator control should not be a shot in the dark. *Front. Ecol. Environ.* 14, 380–388.
- Treves, A., Artelle, K.A., Darimont, C.T., Parsons, D.R., 2017. Mismeasured mortality: correcting estimates of wolf poaching in the United States. *J. Mammal.* 98, 1256–1264.
- Treves, A., Krofel, M., Ohrens, O., Van Eeden, L.M., 2019. Predator control needs a standard of unbiased randomized experiments with cross-over design. *Front. in Ecol. Evol.* 7, 402–413.
- Wielgus, R.B., Peebles, K., 2014. Effects of wolf mortality on livestock depredations. *PLoS One* 9, e113505.
- Wilkinson, C.E., McInturff, A., Miller, J.R., Yovovich, V., Gaynor, K.M., Calhoun, K., Karandikar, H., Martin, J.V., Parker-Shames, P., Shawler, A., Van Scoyoc, A., 2020. An ecological framework for contextualizing carnivore–livestock conflict. *Conserv. Biol.* 34, 854–867.
- Williams, C.L., Blejwas, K., Johnston, J.J., Jaeger, M.M., 2003. A coyote in sheep's clothing: predator identification from saliva. *Wildl. Soc. Bull.* 31, 926–932.
- Woodroffe, R., Ginsberg, J.R., 1998. Edge effects and the extinction of populations inside protected areas. *Science* 280, 2126–2128.
- Woodroffe, R., Donnelly, C.A., Jenkins, H.E., Johnston, W.T., Cox, D.R., Bourne, F.J., Cheeseman, C.L., Delahay, R.J., Clifton-Hadley, R.S., Gettinby, G., Gilks, P.J., Hewinson, R.G., Mcinerney, J.P., Morrison, W.I., 2006. Culling and cattle controls influence tuberculosis risk for badgers. *Proc. Natl. Acad. Sci. U. S. A.* 103, 14713–14717.
- Wydeven, A.P., Mladenoff, D.J., Sickley, T.A., Kohn, B.E., Thiel, R.P., Hansen, J.L., 2001. Road density as a factor in habitat selection by wolves and other carnivores in the Great Lakes region. In: *Endangered Species Update*, 18, pp. 110–114.
- Wydeven, A.P., Treves, A., Brost, B., Wiedenhoef, J.E., 2004. Characteristics of wolf packs in Wisconsin: identification of traits influencing depredation. In: Fascione, N., Delach, A., Smith, M.E. (Eds.), *People and Predators: From Conflict to Coexistence*. Island Press, Washington, D. C., pp. 28–50.
- Young, J.K., Draper, J., Breck, S., 2019. Mind the gap: experimental tests to improve efficacy of fladry for nonlethal management of coyotes. *Wildl. Soc. Bull.* 43, 1–7.
- Zimmermann, B., Sand, H., Wabakken, P., Liberg, O., Andreassen, H.P., 2015. Predator-dependent functional response in wolves: from food limitation to surplus killing. *J. Anim. Ecol.* 84, 102–112.